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Original Research Article

**APPLICATION OF QUALITY INDICATORS IN
THE EVALUATION OF SUBTROPICAL
ARGIUDOLLS AND HAPLUDOLLS IN
FORMOSA (ARGENTINA)**

ABSTRACT

The quality indicators are suitable tools to determine the state of the soil and the effects of different uses and management on it. The aim of the present work was to evaluate the quality of sub-tropical Argiudolls and Hapludolls subjected to different uses in Formosa, using a minimum set of indicators (MSI). Changes in soil use and the application of management techniques to maximize agricultural production are frequent in the world. In Formosa, Argentina, improvements in the productive infrastructure and low market value of the land, promote that these changes occur faster than the monitoring of the ones. The effects of 25 years of continuous agricultural use, extensive livestock in implanted pastures and fruit crops, in relation to the native forest were analyzed. The MSI consisted of five variables: total organic carbon, particulate organic carbon, total nitrogen, structural stability and bulk density. The baseline of the indicators was determined and threshold values were established. The standardized MSI was analyzed graphically. Particulate carbon and structural stability were the most sensitive indicators. Continuous agriculture degraded the edaphic system, resulting in lower values of indicators than the thresholds. It produced a decrease of 74% of the particulate organic carbon and 63% of structural stability, with possible impact on the resilience of the system. Fruit crops led to a decrease in soil quality causing particulate organic carbon and structural stability to approach values that compromise their natural recovery. The

implanted pasture improved the quality of the soil with respect to the degraded native forest.

Keywords: application, soil quality indicators, graphical analysis

1. INTRODUCTION

Changes in soil use and the application of demand-driven management techniques to maximize agricultural production are becoming more frequent in various regions of the world. Some authors have named agriculturization the process that leads to the replacement of crops and conventional uses of the soil, by annual species [1, 2, 3]. The market value of the land in the province of Formosa, Argentina, is lower than other regions; but due to advances in its productive infrastructure, such as road network and electricity, these land use changes occur faster than the alterations produced on the soil that can be monitored. They affect the quality or health of soils and have been reported in different situations [3, 4].

Ferreras et al. [5]; Rojas et al. [6]; Bravo-Medina et al. [7]; Rojas et al. [8], among others, agree in pointing out some quality indicators as tools to assess the state of the soil and the effects of different uses and management of it. An indicator is a variable which summarizes or simplifies relevant information so that the phenomenon or condition of interest can be perceived. It quantifies measures and makes that information available in a comprehensible way, Cantú et al. [9]. Quality indicators are those properties and processes of the soil with greater sensitivity to changes in their functions [10].

For more than 15 years it has been discussed which indicators are suitable for different regions, soils and agroecosystems. Its effective usefulness to monitor soil quality demands basic information, complementary to the identification of the indicators. Arshad and Martín [11], point out that different agro-ecological regions require the selection of a set of indicators and the establishment of their critical limits (threshold values). Segnestam [12] and Wilson and Sione [13] highlight the importance of establishing the reference baseline or starting point for each indicator, in order to reflect the generation of positive or negative impacts on the environment, together with the threshold values (TV) for the monitoring of negative impacts, which must not exceed a certain level or value. This allows to monitor the normal functioning of the soil, to detect changes and to determine the trends of improvement or deterioration in its quality.

The soil indicators interact with each other and, therefore, the value of one is affected by one or more of the selected parameters. A minimum set of indicators (MSI) is a small group of variables capable of synthesizing most of the total variability of a soil in a given site [14].

Some researchers have proposed procedures to evaluate soil quality functions by combining and integrating specific elements in the soil quality indices. These

53 procedures allow us to weigh several functions, depending on the user's objectives
54 and socioeconomic concerns [11]. The indicators are measured in different units, so
55 it is necessary to transform the values of laboratory or field to a single scale,
56 standardizing the results in order to integrate them, for example in indices or graphic
57 representations [6]. Recent applications of indicators use methodologies based on
58 scoring systems to standardize the values of the variables [7, 8] and then integrate
59 them mathematically, statistically and / or graphically for analysis.
60 The objective of the present work was to evaluate the quality of Argiudolls, and
61 subtropical Hapludolls subjected to different uses, using a minimum set of indicators.
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63 2. MATERIALS AND METHODS.

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65 We worked in an area of 24,000 ha, with an approximate centre at 25° 00' South
66 latitude, 58° 30' West longitude, in the departments of Pilagás and Pilcomayo,
67 province of Formosa, Argentina, Figure 1.

68 Data obtained by Baridón and Casas [3] on Argiudolls and Hapludolls [15], of the
69 marginal fluvial hill of El Porteño river, subject to different uses were used. Twelve
70 sampling sites were incorporated following the technique used by these authors.
71 Through satellite information, background and fieldwork, plots with 20 to 25 years
72 of continuous agricultural use in conventional tillage (CA), extensive livestock in
73 implanted pastures (P) and fruit crops (FC), as well as areas with degraded native
74 forest were located (NF). In the first three uses the native forest was clearcut, with
75 mechanical system, using bulldozer, extraction of few individuals for timber
76 purposes, formation of cordons with forest stubble and burning. Then the stems and
77 woody roots are removed by hand and two or more heavy disc harrow passes.

78 The general characteristics of these uses are the following:

79 - Continuous agriculture in conventional tillage (CA): Areas disassembled 25 years
80 ago (± 2). The historical crops were corn (*Zea mays*) and cotton (*Gossypium spp*); in
81 the last 10 years they have decreased and soybean (*Glycine max*) has increased. The
82 tillage system is conventional, heavy plowing is often used.

83 - Pasture (P): Areas disassembled 25 years ago (± 2) where monospecific pastures
84 have predominated, used during average periods of 4 years. The most frequent
85 species are the Dicantium (*Dichanthium spp.*).

86 - Fruit crops (FC). There are mainly two plurianual crops: grapefruit (*Citrus*
87 *paradisi*) and banana (*Musa spp*). The most frequent is the plantation of more than
88 20 years (± 2) in both crops. There are commercial plantations of a younger age (10
89 years old) managed more intensively.

90 - Degraded native forest (NF): forest of degraded native species by grazing, opening
91 of bites and thinnings of wood. The dominant trees are: quebracho colorado
92 (*Schinopsis balansae*), quebracho blanco (*Aspidosperma quebracho-blanco*),
93 guayacán (*Caesalpinia paraguariensis*), urunday (*Astronium balansae*), white carob
94 (*Prosopis alba*) and black carob (*Prosopis nigra*), among others.
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Figure 1. Work area location.



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100 Soil sampling was carried out according to a stratified random design. The strata
101 represented the three situations of use: CA; P; FC and the situation considered as a
102 blank, NF, because it almost represents all the forests present in the study area.
103 Thirty sampling points were randomly distributed on each stratum. A sample from 0
104 to 10 cm deep, composed of 3 subsamples was taken on each sampling point. A total
105 of 120 soil samples were analyzed

106 In order to determine the quality of the soil resulting from the different uses, a MSI
107 was used, composed of five variables: total organic carbon (TOC), particulate
108 organic carbon (POC), total nitrogen (TN), stability structural (SS) and bulk density
109 (BD). The quality indicators of the studied soils indicated by Baridón and Casas [3]
110 were TOC, POC and SS, accompanied by dehydrogenase activity. The latter of a
111 more difficult determination was not used in this work. Both BD and TN have been
112 used in the region as quality indicators [6] and have become routine determinations
113 for technicians and laboratories.

114 The analytical methods used were: total organic carbon by Walkley Black;
115 particulate organic carbon by physical fractionation [16]; Total nitrogen by Kjeldahl
116 method; structural stability by Le Bissonnais method [17], and bulk density by the
117 cylinder method [18].

118 The base line of the indicators, defined as "initial values" (IV), was established from
 119 the average values corresponding to a relic of native forest located within the
 120 Pilcomayo National Park, Formosa province, 25°04' South latitude, 58°07' West
 121 longitude. The TV were set based on bibliographic background and consultations
 122 with experts.

123 The laboratory results were characterized by descriptive statistics. The analysis of
 124 the variance and the comparison of means were carried out using Fisher's minimum
 125 significant difference test.

126 A systematization of the indicators was carried out: the data were normalized
 127 according to a scale of 0 to 8. The scale was estimated individually, for each
 128 indicator, considering optimal and critical reference levels. Table 1 presents the
 129 scales used for normalization.

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131 **Table 1.** Analytical data standardization criteria.

TOC [g.kg ⁻¹]		POC [g.kg ⁻¹]		SS [WAD in mm]		BD [g.cm ⁻³]		TN [g.kg ⁻¹]	
<2	0	<1	0	<0.9	0	> 1.7	0	<0.5	0
2 - 3.9	1	1 - 3	1	0.9 - 1.1	1	1.61 - 1.70	1	0.5 - 0.7	1
4 - 6	2	3.1 - 5	2	1.2 - 1.4	2	1.51 - 1.60	2	0.8 - 1.0	2
6.1 - 10	3	5.1 - 7	3	1.5 - 1.7	3	1.41 - 1.50	3	1.1 - 1.3	3
10.1 - 15	4	7.1 - 9	4	1.8 - 2.0	4	1.31 - 1.40	4	1.4 - 1.6	4
15.1 - 20	5	9.1 - 11	5	2.1 - 2.3	5	1.21 - 1.30	5	1.7 - 1.9	5
20.1 - 25	6	11.1 a 13	6	2.4 a 2.6	6	1.11 a 1.20	6	2.0 a 2.4	6
25.1 - 30.4	7	13.1 a 15	7	2.7 a 2.9	7	1 a 1.10	7	2.5 a 2.9	7
>30.4	8	>15	8	> 2.9	8	<1	8	>2.9	8

132 TOC: total organic carbon; POC: particulate organic carbon; SS: structural stability; BD:
 133 bulk density; TN: total nitrogen; WAD: weighted average diameter

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135 The assessment of the soil quality in the different uses was made by using a star
 136 diagram together with reference values and threshold for the MSI.

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139 **3. RESULTS**

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141 The mean values of the indicators, the variance and the existence of statistically
 142 significant differences between the different land uses are presented in Table 2,
 143 where it is observed that all the indicators show statistical differences between the
 144 different land uses.

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Table 2. Quality indicators, statistical summary.

Quality indicators										
Soil use	TOC [g.Kg ⁻¹]		POC [g.Kg ⁻¹]		SS-WAD [mm]		BD [g.cm ⁻³]		TN [g.Kg ⁻¹]	
	Mean	S	Mean	S	Mean	S	Mean	S	Mean	S
NF	27.04c	6.04	9.31d	3.15	1.98c	0.42	1.17a	0.10	2.21b	0.57
P	32.50d	6.42	7.32c	1.90	2.29c	0.51	1.27b	0.09	3.12c	0.76
FC	22.59b	3.46	4.52b	1.34	0.97ab	0.31	1.23ab	0.07	1.90a	0.37
CA	18.03a	5.16	2.30a	0.93	0.71a	0.25	1.36c	0.09	1.81a	0.45

147 TOC: total organic carbon; POC: particulate organic carbon; SS: structural stability; WAD:
 148 weighted average diameter; BD: bulk density; TN: total nitrogen; S: standard deviation.
 149 Different letters indicate statistical differences between means (p= 0.05)

151 Table 3 shows the IV and the TV of the MSI.

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Table 3. Reference values of soil quality indicators.

INDICATOR	UNIT	Analytical method	Initial values (IV)	Threshold values (TV)
TOC	g.Kg ⁻¹	Walkley Black	39,5	6 (min) †
POC	g.Kg ⁻¹	Galantini (2005)	18,6	4 (min) ‡
SS (WAD)	mm	Le Bissonnais (1996)	2,5	1 ‡
TN	g.Kg ⁻¹	Kjeldahl	2,8	1 ‡
BD	g.cm ⁻³	Cylinder method (Forsythe 1975)	1,1	1,6 (max) ‡

154 † Minimum value of TOC established as a requirement for a mollic epipedon [15].

155 ‡ Background and consultation with experts. TOC: total organic carbon; POC: particulate
 156 organic carbon; SS: structural stability; WAD: weighted average diameter; BD: bulk density;
 157 TN: total nitrogen.

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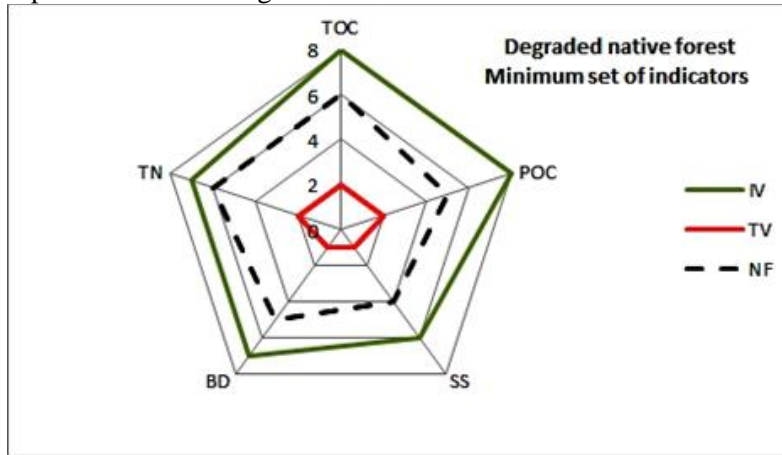
In the present work, these IVs represent the pristine situation of the native forest on a marginal fluvial hill, obtained in the Pilcomayo National Park. They are considered as "base line", even though the blank was a degraded native forest, because these are currently the almost unique expression of the high forest in the area of work. The TV establish the limit values, which, if exceeded by each indicator, compromise the resilience of the system.

Figures 2, 3, 4 and 5 shows, through star diagrams, the comparison between the MSI in the blank, NF; the IV of pristine native forest and the TV; the contrast between the

167 MSI in different land uses and the current state of the indicators in the NF and the
 168 location of each indicator, in each land use, with reference to IV and TV.

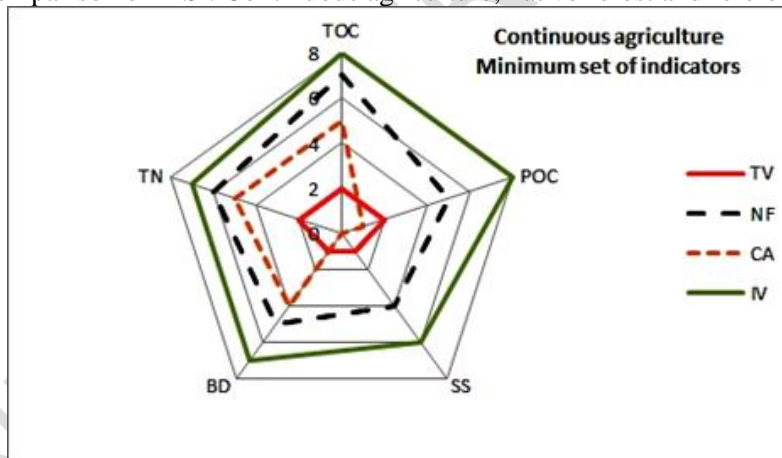
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171 **Figure 2.** Comparison of MSI. Degraded native forest in relation to reference values.
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173 TOC: total organic carbon; POC: particulate organic carbon; SS: structural stability; BD:
 174 bulk density; TN: total nitrogen; IV: initial values; TV: threshold values; NF: native forest.
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176 **Figure 3.** Comparison of MSI. Continuous agriculture, native forest and reference values.
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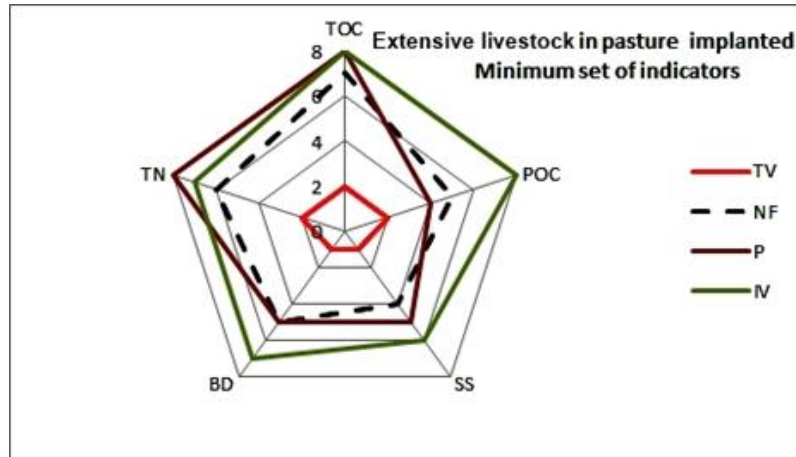


178 TOC: total organic carbon; POC: particulate organic carbon; SS: structural stability; BD:
 179 bulk density; TN: total nitrogen; TV: threshold values; NT: native forest; CA: continuous
 180 agriculture; IV: initial values.

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187 **Figure 4.** Comparison of MSI. Extensive livestock in pasture implanted, native forest and
 188 reference values.

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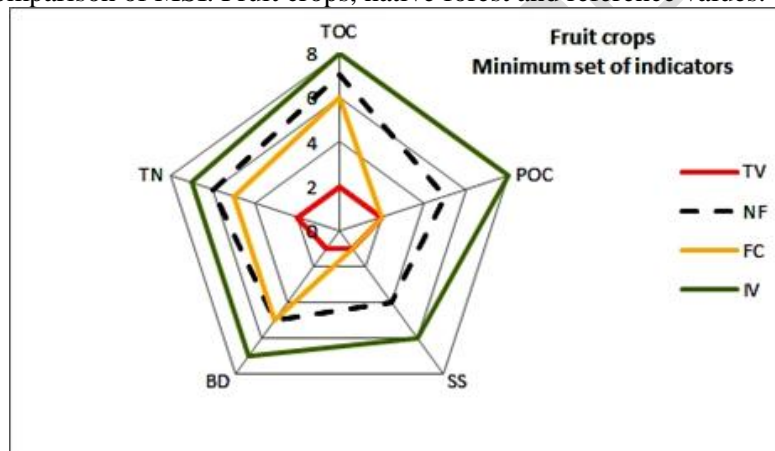


190 TOC: total organic carbon; POC: particulate organic carbon; SS: structural stability; BD:
191 bulk density; TN: total nitrogen; TV: threshold values; NF: native forest; P: pasture; IV:
192 initial values.

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194 **Figure 5.** Comparison of MSI. Fruit crops, native forest and reference values.

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196 TOC: total organic carbon; POC: particulate organic carbon; SS: structural stability; BD:
197 bulk density; TN: total nitrogen; TV: threshold values; NF: native forest; FC: fruit crops; IV:
198 initial values.

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200 4. DISCUSSION

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202 A first analysis of Table 2 data indicates that continuous agriculture is the cause of
203 the greatest loss of soil quality. In a similar way, although to a lesser extent, the fruit
204 crops also produced negative effects in all the parameters evaluated. The extensive
205 livestock with implanted pasture caused an increase in the total carbon content, total
206 nitrogen and stability of the aggregates, with respect to the average values of these
207 parameters in the native degraded forest. The lowest values of standard deviation
208 were registered in almost all cases under fruit crops (Table 2), possibly due to the
209 uniformity in the management of this productive system.

210 The joint presentation of the standardized indicators (Figures 2, 3, 4 and 5) allows
211 the comparative evaluation of soil quality after 25 years of different uses. In almost
212 all cases the separation of the MSI from the initial values (IV) is visualized, some of
213 them exceeding the threshold values (TV), which could compromise the resilience of
214 the edaphic system [13].

215 Figure 2 shows the degradation of the soil in the native forests that now exist in the
216 study area, as a result of selective interventions for the extraction of wood and
217 firewood, together with grazing, mainly of goats and cattle. From the forestry point
218 of view, the forests have reduced their specific heterogeneity, reduced their canopy,
219 lost individuals from the upper strata and increased the density of plants in the lower
220 strata. The natural regeneration of the forest has been compromised. The
221 contributions of organic matter to the soil from natural forest, heterogeneous and
222 with high nutrient contents have been modified. Prause et al. [19], when evaluating
223 contributions of dry matter and composition of the litter of Quebracho colorado
224 (*Schinopsis balansae*), reported contributions of $1.36 \text{ Mg}\cdot\text{ha}^{-1}\cdot\text{year}^{-1}$ of dry matter.
225 This litter, containing $11.7 \text{ mg}\cdot\text{g}^{-1}$ of nitrogen and $60 \text{ g}\cdot\text{kg}^{-1}$ of phosphorus, for
226 example, is practically not present in the NF, since the individuals of Quebracho
227 colorado have been the most extracted. The decrease in TOC, POC and TN directly,
228 as well as the loss of SS and the increase in BD associated with the deterioration of
229 the porous soil system in the topsoil, are a consequence of unsustainable use of the
230 forest resource.

231 After 25 years of continuous agriculture, in conventional sowing, the deterioration of
232 soil quality has increased, as shown in the graph of the MSI, Figure 3. Cantú et al.
233 [9]; Piccolo et al. [20]; Ferreras et al. [5]; Moges et al. [21]; among others, have
234 evaluated the organic carbon content as an indicator in different soils and regions. In
235 the study area, the continuous realization of conventional tillage in CA use has
236 favoured the mineralization of organic matter, which would justify the decrease of
237 29% in the content of TOC with respect to NF. This loss is somewhat lower than that
238 reported by Moges et al. [21], for subtropical soils of Ethiopia and similar to that
239 reported by Ferreras et al. [5] in Argiudolls in the provinces of Santa Fe and
240 Córdoba, Argentina. Figure 3 shows that the values of POC and SS have exceeded
241 the threshold values. The POC has decreased 74% with respect to the NF, and 87%
242 with respect to the pristine native forest; thus also the SS has been reduced 63% with
243 respect to NF degraded, and 70% with respect to the pristine forest. TOC, TN and
244 BD also point to a loss of soil quality; however they are still in the range of values in
245 which the degradation process could be reversed. The implementation of direct
246 sowing and the use of crops to maximize the amount of surface residues are frequent
247 practices in order to increase the organic carbon content in the soil [22]. Céspedes et
248 al. [23], working in a drier area of the Chaco region, reported that the highest
249 contributions of surface carbon were produced by meadows, and by the forest in the
250 first 15 cm of the soil. The rotation of crops allows to improve the soil carbon
251 balance [24]; rotations that include pastures and service crops should be applied
252 immediately in this situation of use.

253 The MSI in the fruit crops, (Figure 5) indicates a loss of soil quality. The decrease
254 with respect to NF in the POC content is highlighted, 48%, and 47% of SS; in both
255 indicators reaching values close to the thresholds. Ferreras et al. [5] and de
256 Figueiredo et al. [25], agree in pointing to the particulate fraction of organic carbon
257 as the one where the greatest changes occur in the face of different managements.
258 After the clearing and preparation of the soil for the implantation of the fruit tree,
259 soil tillage is generally limited to weed control by means of disc harrowing between
260 rows during the first years. This would have caused the decrease of 15% of TOC
261 with respect to NF; the TN has accompanied this trend. Although there were
262 statistical differences in BD values, they do not seem agronomically relevant.
263 Contrary to other land uses, the implanted pasture, (Figure 4), has achieved an
264 improvement in the TOC, TN, and SS indicators, recomposing the blank values, NF
265 degraded, and even exceeding the content of TN to the pristine native forest. The
266 increase in the content of TN for use P, 30.6% on degraded NF and 9.6% on the
267 initial values, is related to the quantity, quality and distribution of the organic matter
268 that the prairies incorporate into the soil every year. The content of TN in NF (2.35
269 g.Kg^{-1}) is coincident with the behavior of TN under scrub vegetation, even in
270 conditions of lower humidity, reported by Albanesi et al. [26] in Argiustolls and
271 Haplustolls of the semiarid Chaco region. The IV of TN corresponding to the pristine
272 situation of the native forest are related to the high contributions by leaf litter of
273 species such as *Schinopsis balansae*, which are around 11.7 mg. g^{-1} of N in the
274 humid Chaco region [19].
275 The decrease of the content of TN in CA and FC situations would be associated with
276 the elimination of superficial microbial crusts composed of cryptogams,
277 cyanobacteria, lichens and microscopic fungi, present in the first centimeters of the
278 forest floor [26] since they constitute a dynamic source of nitrogen in the Chaco
279 region.
280 The POC was a sensitive indicator in all evaluated soil uses confirming its ability to
281 respond to different management reported by other authors [5, 25]. Duval et al. [27]
282 confirm that the labile forms of carbon are sensitive to soil management: they
283 obtained sensitive indicators in short periods of time and direct sowing conditions
284 when separating the fractions of coarse particulate carbon and fine particulate
285 carbon.
286 As the graphical analysis of the MSI indicates (Figure 2) the use of continuous
287 agriculture with conventional tillage and fruit monoculture has produced the
288 degradation of soil quality.
289 Wilson and Sione [13] point out that when quantifying the deterioration of the
290 edaphic qualities it is essential to establish the reference base line or starting point
291 for each indicator, to reflect the generation of positive or negative impacts. Based on
292 the results obtained, it is necessary to reconsider some of the threshold values used.
293 The required TOC content for the mollic epipedon, 6 g.kg^{-1} [15], for example, is a
294 very low threshold if we want to consider that degradation compromises the
295 resilience of the system.
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5. CONCLUSION

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The use of continuous agriculture with conventional tillage and fruit monoculture has produced the degradation of soil quality in Argiudolls, and subtropical Hapludolls. In

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both cases all indicators accused the degradation.

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Continuous agriculture in conventional sowing degraded the edaphic system with possible affectation of its resilience, associated to the loss of labile forms of carbon and to the decrease of the structural stability, with these indicators below the critical levels.

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Fruit crops use led to a decrease in soil quality with the particulate carbon and the stability of the aggregates to approach values that compromise their natural recovery.

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The implanted pasture favored the improvement in the quality of the soil with respect to the degraded native forest, however the contents of the light carbon fractions continued below those of the degraded native forest..

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The MSI used and its statistical and graphical analysis allowed to evaluate the quality of Hapludolls and Argiudolls against 25 years of agricultural, fruit and pastoral use.

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COMPETING INTERESTS

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Authors have declared that no competing interests exist.

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AUTHORS' CONTRIBUTIONS

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This work was done in collaboration among the two authors. Both authors read and approved the final manuscript.

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REFERENCES

326

327

1. Duhour A, Costa C, Momo F, Falco L, Malacalza L. (2009). Response of earthworm communities to soil disturbance: Fractal dimension of soil and species' rank-abundance curves. *Applied Soil Ecology*, 43, 83-88. Doi:10.1016/j.apsoil.2009.06.004

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2. Manuel-Navarrete D, Gallopín G, Blanco M, Díaz-Zorita M, Ferraro D, Herzer H, Lathera P, Murmis M, Podestá G, Rabinovich J, Satorre E, Torres F, Viglizzo E. (2009). Multi-causal and integrated assessment of sustainability: the case of agriculturization in the Argentine Pampas. *Environment, Development and Sustainability*, 11(3), 621-638. Doi: 10.1007/s10668-007-9133-0

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337

3. Baridón J.E, Casas R.R. (2014). Quality indicators in subtropical soils of Formosa, Argentina: Changes for agriculturization. *International Soil and Water Conservation Research*, 2 (4), 13-24. Doi: 10.1016 / S2095-6339 (15) 30054-X

338

339

340

4. Baridón E, Pellegrini A, Lanfranco J, Cattani V. (2012). Variation of the organic fraction by agriculturalization in subtropical Alfisoles of Argentina. *Agri Sciences*, 2 (3), 371-378.

341

- 342 5. Ferreras L, Toresani S, Bonel B, Fernández E, Bacigaluppo S, Faggioli V, Beltrán
343 C. (2009). Chemical and biological parameters as indicators of soil quality in
344 different management. *CI. Soil (Argentina)*, 27 (1), 103-114.
- 345 6. Rojas JM, Prause J, Sanzano G, Ernesto O, Arce A, Sánchez MC. (2016). Soil
346 quality indicators selection by mixed models and multivariate techniques in
347 deforested areas for agricultural use in NW of Chaco, Argentina. *Soil & Tillage
348 Research*, 155, 250-262. Doi: 10.1016 / j.still. 2015.08.010
- 349 7. Bravo-Medina C, Marín H, Manero-Labrador P, Ruíz M, Torres-Navarrete B,
350 Navarrete-Alvarado H, Durazno-Alvarado G, Changoluisa-Vargas D. (2017).
351 Sustainability evaluation through indicators in production units in the province of
352 Napo, Ecuadorian Amazonia. *Bioagro*, 29 (1), 23-36.
- 353 8. Rojas J, Mórtoles N, Romaniuk R, Russo E. (2018). Guide for assessing the quality
354 of soils under agriculture in Sub-humid Chaco. National Institute of Agricultural
355 Technology (INTA). Available from:
356 https://inta.gob.ar/sites/default/files/guia_evaluacion_calidad_de_suelos_chaco.pdf
- 357 9. Cantú MP, Becker A, Bedano JC. (2007). Evaluation of soil quality through the
358 use of indicators and indices. *CI. Soil (Argentina)*, 25 (2), 173-178.
- 359 10. Giuffré L, Ratto S, Romaniuk R. (2008). Environmental indicators. In:
360 *Agrosystems: Environmental Impact and Sustainability*. Lidia Giffre, Editor.
361 Editorial Faculty of Agronomy. University of Buenos Aires, 1-17.
- 362 11. Arshad M, Martín S. (2002). Identifying critical limits for soil quality indicators
363 in agro-ecosystems. *Agriculture, Ecosystems & Environment*, 88 (2), 153-160.
- 364 12. Segnestam L. (2002). Indicators of the environmental and sustainable
365 development. *Theories and Practical Experience*. Environmental Economic Series,
366 Paper No. 89, World Bank, Washington D.C. 61 p.
- 367 13. Wilson M.G, Sione S.M. (2017). Selection of soil quality indicators. Obtaining
368 the minimum set of indicators (CMI). Criteria for defining threshold values and
369 obtaining indexes. 93-105p. In: Wilson M.G, editor. "Manual of soil quality
370 indicators for the ecoregions of Argentina". Digital book. 2017. National Institute of
371 Agricultural Technology (INTA). Available from:
372 https://inta.gob.ar/sites/default/files/manual_ics_final.pdf.
- 373 14. Rezaei S.A, Gilkes R.J, Andrews S.S. (2006). A minimum data set for assessing
374 soil quality in rangelands. *Geoderma*, 136(1-2), 229-234. Doi:
375 10.1016/j.geoderma.2006.03.021
- 376 15. Soil Survey Staff. (2014). *Keys to Soil Taxonomy*. 12th ed. Washington, DC:
377 USDA-Natural Resources Conservation Service. 399 p.
- 378 16. Galantini J.A. (2005). Separation and analysis of organic fractions. In: Marban L,
379 Ratto S, editors. *Manual "Technology in Soil Analysis: Reaches to agricultural
380 laboratories"* of the AACS. Chapter IV part 2, 103-114.
- 381 17. Le Bissonnais YL. (nineteen ninety six). Aggregate stability and assessment of
382 soil crustability and erodibility: I. Theory and methodology. *European Journal of
383 Soil Science*, 47 (4), 425-437.
- 384 18. Forsythe W. (1975). *Laboratory manual: soil physics*. San José, Costa Rica:
385 IICA. 212 p.

- 386 19. Prause J, Fernández López C, Contreras Leiva S.M, Gallardo Lancho JF. (2012).
387 Contribution and decomposition of leaves and re-absorption of N, P and K in a
388 primary forest of *Schinopsis balansae* Engler with and without silvopastoral
389 management in the humid Chaco park. *FACENA*, 28, 41-50. Doi: 10.30972 /
390 fac.280900
- 391 20. Piccolo GA, Andriulo AE, Mary B. (2008). Changes in soil organic matter under
392 different land management in Misiones Province (Argentina). *Sci. Agric.*
393 (Piracicaba, Braz.), 65 (3), 290-297. Doi: 10.1590 / S0103-90162008000300009
- 394 21. Moges A, Dagnachew M, Yimer F. (2013). Land use effects on soil quality
395 indicators: A case study of Abo-Wonsho Southern Ethiopia. *Applied and*
396 *Environmental Soil Science*. Available from: <https://doi.org/10.1155/2013/784989>
- 397 22. Galantini J, Iglesias J, Landriscini M.R, Suñer L, Minoldo G. (2008). Quality and
398 dynamics of organic fractions in natural and cultivated systems. In: Galantini J,
399 editor. *Study of organic fractions in soils of Argentina*. Bahía Blanca, Argentina:
400 EdiUNS, 71-95.
- 401 23. Céspedes Flores F.E, Fernández J, Giménez L, Leonhardt E, Bernardis C. (2018).
402 Carbon retained by litter and roots in different land uses in the semi-arid Chaco
403 region. *Chilean Journal Agric. Anim Sci. (Former Agro-Science)*, 34 (2), 165-172.
404 Doi: 10.4067 / S0719-38902018005000405.
- 405 24. Andriulo A, Sasal MC, Irizar A, Restovich S, Rimatori F. (2008). Effects of
406 different tillage systems, crop sequence and nitrogen fertilization on edaphic C and N
407 stocks. In: Galantini J, editor. *Study of organic fractions in soils of Argentina*. Bahia
408 Blanca, Argentina: UNS. 117-129.
- 409 25. de Figueiredo CC, Siqueira Resck DV, Carbone Carneiro MA. (2010). Labile
410 and stable fractions of soil organic matter under management systems and native
411 Closed. *R. Bras. Ci. Alone*, 34, 907-916. Doi: 10.1590 / S0100-06832010000300032
- 412 26. Albanesi A, Anríquez A, Polo Sánchez A. (2001). Effects of conventional
413 agriculture on some forms of nitrogen in a toposequence of the Chaco region,
414 Argentina. *Agriscientia*, 18, 3-11.
- 415 27. Duval M.E, Galantini J.A, Martínez J.M, López F.M, Wall L.G. (2016).
416 Sensitivity of different soil quality indicators to assess sustainable land management:
417 Influence of site features and seasonality. *Soil and Tillage Research*, 159, 9-22. Doi:
418 10.1016 / j.still. 2016.01.004